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Evidence does not support a role for gallic acid in *Phragmites australis* invasion success

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Abstract—Gallic acid has been reported to be responsible for the invasive success of nonnative genotypes of *Phragmites australis* in North America. We have been unable to confirm previous reports of persistent high concentrations of gallic acid in the rhizosphere of invasive *P. australis*, and of high concentrations of gallic acid and gallotannins in *P. australis* rhizomes. The half-life of gallic acid in nonsterile *P. australis* soil was measured by aqueous extraction of soils and found to be less than 1 day at added concentrations up to 10,000 $\mu\text{g g}^{-1}$. Furthermore, extraction of *P. australis* soil collected in North Carolina showed no evidence of gallic acid, and extractions of both rhizomes and leaves of samples of four *P. australis* populations confirmed to be of invasive genotype show only trace amounts of gallic acid and/or gallotannins. The detection limits were less than 20 $\mu\text{g gallic acid g}^{-1}$ FW in the rhizome samples tested, which is approximately 0.015% of the minimum amount of gallic acid expected based on previous reports. While the occurrence of high concentrations of gallic acid and gallotannins in some local populations of *P. australis* cannot be ruled out, our results indicate that exudation of gallic acid by *P. australis* cannot be a primary, general explanation for the invasive success of this species in North America.

Key Words—*Phragmites australis*, allelopathy, gallic acid, invasive species, novel weapons hypothesis.

INTRODUCTION

Phragmites australis (Cav.) Trin. ex Steud., a common marsh grass in wetlands in the U.S., has become a serious invasive species. Its abundance has increased greatly over the past 150-200 years, and Saltonstall (2002) found that invasive populations of *P. australis* possessed a non-

native, Eurasian genotype and that these plants had displaced those of a native North American lineage, especially in New England and along the Atlantic Coast.

It has recently been reported (Rudrappa et al. 2007) that gallic acid is released and maintained in extremely high concentrations ($>10,000 \mu\text{g g}^{-1}$ soil) beneath invasive *P. australis* genotypes and may be responsible for the invasive success of this plant. Rudrappa et al. (2007) found that root exudates of invasive *P. australis* plants were more toxic than native genotypes to test plants, and gallic acid (3, 4, 5-trihydroxybenzoic acid) was identified as the principal phytotoxin of the root exudates. Gallic acid was found to trigger a “suicidal cell death cascade” in test plants through the generation of toxic levels of reactive oxygen species (Rudrappa and Bais 2008). In subsequent work, Bains et al. (2009) reported that rhizomes of invasive *P. australis* plants contained elevated levels of gallotannins along with free gallic acid, and that native plants and microorganisms produced high levels of tannase that resulted in the release of high amounts of gallic acid in the root zone of *P. australis* by hydrolysis of these gallotannins. Bains et al. (2009) proposed that this represents a previously unobserved extension of the novel weapons hypothesis (Callaway and Ridenour 2004), whereby neighboring plants and native microorganisms activate a less toxic precursor exuded by roots of an exotic invader to the detriment of the invaded community. Rudrappa et al. (2009) reported that ultraviolet light resulted in the photo-degradation of gallic acid in the *P. australis* rhizosphere and that the degradation product mesoxalic acid was toxic to seedlings through a mechanism similar to gallic acid.

Gallic acid has previously been reported as one of the compounds responsible for the allelopathic effects of the perennial shrub *Polygonella myriophylla* (Small) Horton, which grows in the

Florida sand pine scrub community (Weidenhamer and Romeo 1989, 2004). *P. myriophylla* is one of several fire-sensitive shrubs in this community that use allelopathic mechanisms to deter invasion of native and exotic grasses and herbs that could provide fuel for fires (Weidenhamer and Romeo 2005). Previous work has shown that when gallic acid was added to Florida scrub soils at concentrations as high as 400 $\mu\text{g g}^{-1}$, it completely disappeared within three days in non-sterile soils, but persisted for the length of the study (nine days) in sterile soils (Weidenhamer and Romeo 2004). Other bioassay work established that spiking soil with gallic acid in combination with hydroquinone, the other suspected inhibitor produced by *P. myriophylla*, did cause long-term inhibitory effects on plant growth despite their rapid disappearance, possibly due to the formation of more toxic and persistent breakdown products (Weidenhamer et al. 1989).

The marked difference in the reported persistence of gallic acid in *P. australis* soil compared to what has previously been observed in Florida scrub soils suggested to us that the stability of this phenolic acid may be greatly extended in saturated soils. The initial objective of this study was therefore to characterize the breakdown of gallic acid in sterile and nonsterile *P. australis* soils. Once field-collected soils were found to contain undetectable levels of gallic acid and the rapid breakdown of gallic acid in nonsterile *P. australis* soils was established, a phytochemical investigation of *P. australis* rhizomes was undertaken and both rhizomes and leaves of four populations in Ohio and North Carolina USA were analyzed for gallic acid content.

METHODS AND MATERIALS

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95 *Soil Samples.* Soil was collected from beneath *Phragmites australis* and from an adjacent
96 brackish marsh that did not have any *P. australis* in the Outer Banks of North Carolina, USA
97 (35.8448 N, 75.5660 W). Samples were stored frozen (-20°C) for approximately six months
98 prior to the start of this study. Before use, soils were air-dried, passed through a 2 mm screen to
99 remove roots and other large debris. The soil was then stored frozen until use. A subsample of
100 the *P. australis* soil was analyzed for basic soil parameters by the STAR laboratory (Ohio
101 Agricultural Research and Development Center, Ohio State University) and for soil texture
102 (Table 1).

103

104 *Soil sample preparation.* Initial extractions of the *P. australis* soil described above (thawed and
105 extracted before air drying) found no detectable gallic acid. Therefore, this soil was used for all
106 studies of the behavior of gallic acid in soil. Soil was treated with six concentrations of gallic
107 acid: 250, 500, 1000, 5000, 10,000, and 20,000 µg gallic acid g⁻¹ soil. The upper concentrations
108 used in this study were chosen to be in the range of free gallic acid concentrations reported in *P.*
109 *australis* soil by Rudrappa et al. (2007) and Bains et al. (2009). In order to add these high
110 concentrations of gallic acid to soil, gallic acid was dissolved in absolute ethanol and added in
111 250 µl portions to 25 g soil. The soil was allowed to dry and was mixed well after each portion
112 was added. The total amount of ethanol added was constant for all treatments (4 ml per 25 g
113 soil). Complete evaporation of ethanol was confirmed by weighing. One gram of soil was then
114 weighed into a 15×125 mm test tube, and 0.75 ml of deionized water was added to create a
115 saturated aqueous environment. Finally, test tubes were sealed with Parafilm® and placed in the

dark at ambient laboratory temperature until extraction and analysis. Triplicate samples of each treatment were extracted at 0, 0.5, 1, 1.5, 2, 3 days for rates of 10,000 μg gallic acid g^{-1} soil and below. The initial sampling interval was 0.5 days based on preliminary experiments showing very rapid breakdown of gallic acid in nonsterile soil. Soil treated with 20,000 μg gallic acid g^{-1} was extracted at 0, 1, 2, 3, 5, 7 and 9 days, based on preliminary experiments indicating longer persistence of gallic acid at very high concentrations. The initial analysis (0 days time point) was done immediately after sample preparation. Sterilized soils were prepared in the same way, but autoclaved three times at 48 hr intervals (Dalton et al. 1989). Triplicate samples of each treatment were extracted at 0, 5 and 9 days.

Soil extraction method. The soil extraction method followed that of Weidenhamer and Romeo (2004). Samples were either extracted immediately or frozen to terminate gallic acid breakdown and stored for up to 10 days at -20°C until analysis. Extractions of replicate treated samples extracted immediately and after storage confirmed that concentrations of frozen samples did not change over this length of time. Soil samples were extracted for one hr with 2 ml of distilled water, well mixed by briefly vortexing three times over this period. To insure complete extraction of gallic acid, the extraction volume was increased to 4 ml for sterile soils spiked with concentrations of 5,000 μg g^{-1} and above, and for nonsterile soils spiked with 20,000 μg g^{-1} at 0, 1 and 2 days. After centrifugation, extracts were filtered through a 0.22 μm nylon syringe filter. When necessary, an appropriate dilution was made with water before HPLC analysis.

137 *Genotyping of P. australis.* Samples of *P. australis* were collected from two populations in Ohio
138 (Site 1: 41.1655 N, 81.8590 W; Site 2: 41.1355 N, 81.8058 W) and two populations in North
139 Carolina (Site 1: 35.8448 N, 75.5660 W; Site 2: 36.1696 N, 75.7569 W), USA. Each sample
140 was genotyped according to the restriction fragment-length polymorphism (RFLP) assay
141 published by Saltonstall (2003), which distinguishes between the native North American and
142 invasive Eurasian genotypes. DNA was isolated from young *P. australis* shoots using a 2%
143 CTAB (cetyl tri-methylammonium bromide) extraction procedure modified from Doyle and
144 Doyle (1987).

146 Two regions of chloroplast DNA were amplified using polymerase chain reaction (PCR) with the
147 following primer pairs: *rbcL* forward (5' TGTACAAGCTCGTAACGAAGG) & reverse (5'
148 GATTTGTCAAGTCTCATGATCGT), and *trnL* forward (5' TCTACCGATTTCGCCATATC)
149 & reverse (5' GGAGAAGATAGAATCATAGC). Each PCR reaction mixture consisted of
150 approximately 400 ng of DNA, 7.5 pmol of a forward and reverse primer, 25 µl OneTaq DNA
151 polymerase 2X master mix (New England Biolabs, Ipswich, MA) with standard buffer, and
152 DEPC water up to a final volume of 50 µl. Reactions were amplified in a thermal cycler (Bio-
153 Rad DNA Engine Peltier thermal cycler, Hercules, CA) using the following conditions: 2 min. at
154 94°C, followed by 35 cycles of 94°C for 45 sec., 52°C for 45 sec., and 72°C for 1 min. 30 sec.,
155 and finished by a final extension of 72°C for 2 min. Ten µl of the PCR products amplified with
156 the *rbcL* and *trnLb* primer pairs was digested individually with the enzymes *RsaI* and *HhaI* (New
157 England Biolabs) according to the manufacturer's protocol. The digest products were run on 3%
158 TAE agarose gels with a 50 bp DNA ladder (Fermentas, Glen Burnie, MD). Digestion of *rbcL*

amplification products by *HhaI* are indicative of invasive strains while digestion of *trnLb* amplification products by *RsaI* indicate a native strain (Saltonstall 2003).

P. australis foliar and rhizome extracts. Hydrolyzed extracts of *P. australis* foliage and rhizomes from each of the four populations were prepared according to Harborne (1984) to examine the phenolic aglycones present. Duplicate samples of fresh or frozen leaves or rhizomes (1-3 g fresh weight) were covered with 2 M HCl in a 20 mL vial and placed in a boiling water bath for 30-40 min. After cooling, the aqueous phase was filtered through glass wool and extracted four times with a one fourth volume of ethyl acetate. The ethyl acetate fraction was washed with water, dried with granular anhydrous sodium sulfate, and evaporated with gentle warming in a fume hood. Moisture content of the rhizomes and leaves was determined by drying for 48 hr at 60°C. Extracts were analyzed by HPLC to confirm presence or absence of gallic acid. The retention time of gallic acid was established using an authentic standard (see HPLC analytical methods below). Triplicate samples of gallic acid and methyl gallate (2, 10 and 25 mg) were also subjected to the acid hydrolysis and extraction procedure, followed by HPLC analysis, to verify the success of the mild acid hydrolysis for bound forms of gallic acid as well as the stability and recovery of gallic acid.

A bulk extraction of *P. australis* rhizomes (140 g) using the above methods was undertaken to identify the major phenolic and flavonoid compounds. Initial separation was achieved by preparative thin layer chromatography (TLC) on cellulose using 6% acetic acid as the mobile phase. Further separation was carried out by preparative TLC on silica gel GF using ethyl

181 acetate: benzene (45:55, v/v) or acetic acid: dichloromethane (1:9, v/v) as the mobile phase.
182 Isolated fractions were analyzed by HPLC and by high field NMR. GC-MS analyses were
183 carried out on an Agilent 6890/5975 gas chromatograph/ quadrupole mass spectrometer, using a
184 30 m (250 μ m inner diameter x 0.25 μ m film thickness) HP-5MS column. Starting column
185 temperature was 40°C (3 min hold), and increased at 12°C per min to 260°. The Agilent
186 ChemStation software included the NIST/EPA/NIH Mass Spectral Library (version 2.0). LC-
187 MS analysis was carried out on a Thermo QExactive Orbitrap MS with a Thermo Accela 1250
188 UPLC, using the HPLC column and mobile phase detailed below.

189
190 *HPLC analytical method.* Samples were analyzed using a Restek Ultra C18 column (150 \times
191 4.6mm, 5 μ m particle size) on an Agilent 1100 HPLC system which included an autosampler and
192 diode array detector. Samples were eluted isocratically with water/ methanol/acetic acid
193 (97.9/2.0/0.1, v/v/v) for the first 12 min, followed by gradient to 95% methanol with 5% of water/
194 methanol/acetic acid (97.9/2.0/0.1, v/v/v) over 12-20 min. The mobile phase remained at this
195 composition until 25 minutes, before switching back to the initial conditions. There was a four
196 min equilibration time between runs to allow the system to stabilize. The injection volume was
197 25 μ l, and the detection wavelength was 272nm for gallic acid. All samples were injected in
198 duplicate. Linearity of gallic acid response was established for both peak area vs. concentration
199 of authentic standards ($R^2=0.9998$) and log peak area vs. log concentration ($R^2=0.999$) over a
200 range of 1 to 500 mg l⁻¹. The quantitation limit was conservatively estimated at 0.25 mg l⁻¹, the
201 concentration of the lowest standard used, which corresponds to a concentration in soil of less
202 than 0.5 μ g g⁻¹.

Nuclear Magnetic Resonance Spectra. NMR spectra were obtained on a JEOL ECS 400 MHz spectrometer equipped with a 5mm proton/multi-frequency, autotune and match-ready, z-axis pulsed field gradient, high sensitivity NMR probe. The operating frequencies were 399.782 MHz for ^1H and 100.525 MHz for ^{13}C . Standard JEOL acquisition experiments (^1H , ^{13}C , COSY, HMQC, and HMBC) were used without modification other than varying the number of scans. The spectral assignments were obtained in methanol- d_4 using instrument referencing relative to TMS.

RESULTS

Gallic acid breaks down rapidly in nonsterile P. australis soil. Gallic acid rapidly degraded in nonsterile soil (Figure 1). The half-life of gallic acid increased somewhat with concentration, but remained less than 1 day up to $10,000\ \mu\text{g g}^{-1}$ (Table 2). At $20,000\ \mu\text{g g}^{-1}$ the persistence of gallic acid increased significantly, but increased variability was seen between replicates. Several presumed microbial breakdown products of gallic acid were observed (Figure 2), but these are as yet unidentified. All of the breakdown products eluted prior to gallic acid, indicating that they were more polar than gallic acid. Gallic acid was much more persistent in autoclaved soil (Figure 3), but extractable gallic acid did diminish significantly over time, particularly at the lowest applied rates.

RFLP analysis of Ohio and North Carolina *P. australis* samples indicate that all are invasive strains. Polymorphisms in the chloroplast genome were used to determine whether *P. australis* specimens used in this study were native or invasive. A PCR approach was used to amplify these polymorphic regions and sequences were identified by restriction enzyme digest. PCR amplification of the *rbcL* and *trnLb* regions of our purified *P. australis* DNA resulted in 350 bp fragments as expected (Figure 4). The amplified *rbcL* regions from all four of our collection sites were cut by *HhaI*, but not by *RsaI*. Neither restriction enzyme cut our amplified *trnLb* regions. These results indicate that plant material from all four collection sites are invasive strains of *P. australis* (Saltonstall 2003).

Gallic acid and gallotannins are not major components of *P. australis* leaves or rhizomes. Hydrolysis and extraction of known samples of gallic acid and methyl gallate demonstrated that gallic acid is stable in the extraction procedure, and that methyl gallate (used as a proxy for gallotannins) is effectively hydrolyzed (Table 3). Recovery decreased somewhat with concentration, but reproducibility was good and all recoveries averaged above 70%. Application of this procedure to rhizome and foliage samples from four *P. australis* populations recovered undetectable to trace levels of gallic acid (Table 4). Only three of eight rhizome samples had detectable gallic acid. These peaks were too small to give an acceptable ultraviolet spectrum to compare with gallic acid, so identification was based on peak retention time. None of the rhizome samples showed gallic acid concentrations above the estimated quantifiable detection limit. Gallic acid was detected in seven of eight foliage samples, but again the amounts found were quite low.

246

247 In HPLC monitoring at 272 nm, rhizome samples contained two major components that eluted
248 after gallic acid at approximately 15.9 and 16.5 min (Figure 5), however the relative amounts of
249 these two compounds varied from one population to another and in one case (data not shown)
250 from sample to sample. Monitoring at 360 nm indicated another strongly absorbing component
251 eluting at approximately 19.7 min (data not shown).

252

253 ***Characterization of other P. australis constituents.*** Initial efforts focused on the identification
254 of the major extract component eluting at 15.9 min by HPLC (Figure 5). This compound was
255 isolated from the bulk extraction by preparative TLC and characterized by NMR and mass
256 spectroscopy. ¹H NMR in methanol-d₄ clearly showed evidence of an aldehyde proton (δ = 9.51
257 ppm, s), two aromatic protons (δ = 7.36 ppm, d, J=3.2 Hz; and 6.56 ppm, d, J=3.2 Hz), and a
258 hydroxymethyl (δ = 4.59 ppm, s) substituent. High resolution LC-MS showed a mass of
259 127.0389 (M+1), consistent with a molecular formula of C₆H₆O₃ for the parent compound. The
260 GC-MS spectrum showed major m/z fragments at 126, 97, 69, and 51, consistent with 5-
261 hydroxymethyl-2-furancarboxaldehyde (= 5-HMF, 5-hydroxymethylfurfural, C₆H₆O₃) in the
262 NIST mass spectral database (Supplemental Figure 1). This structure was also consistent with
263 the exact mass obtained by LC-MS. In addition, 2D NMR data (COSY, HMQC and HMBC
264 spectra) supported this structural assignment. The identification was further confirmed by
265 comparison with the ¹H NMR and ¹³C NMR spectra of an authentic sample of 5-HMF (Sigma
266 Chemical Co., St. Louis, MO). GC-MS analysis of the fraction containing 5-HMF also

tentatively identified 2-methoxy-4-vinylphenol as a trace component in the extract
(Supplemental Figure 2).

5-HMF is a common constituent of dried fruits and baked goods, and can be formed as a result of the Maillard reaction occurring when sugars degrade with heating in an acidic environment (Ramírez-Jiménez et al. 2000; Gaspar and Lucena 2009). This suggested the possibility that this compound might be an artifact of the extraction procedure used. To examine this, duplicate 1 g samples of fresh foliage of three species (Sugar maple, *Acer saccharum* Marsh.; corn, *Zea mays* L.; and kohlrabi, *Brassica oleracea* var. *gongylodes*) were carried through this procedure and analyzed by HPLC. All samples were found to contain significant amounts of 5-HMF, verified by both retention time and UV spectrum of the 5-HMF peak. Because of this, our finding of 5-HMF in rhizome and leaf extracts of *P. australis* must be regarded as an artifact of the extraction procedure.

DISCUSSION

Gallic acid degradation in sterile and non-sterile soil. The rapid degradation of gallic acid in non-sterile soil is in agreement with previous work (Weidenhamer and Romeo 2004) and with numerous studies showing that microorganisms are very effective in degrading simple phenolic compounds and other allelochemicals (Schmidt 1988; Blum 1998; Blum et al. 2000; Inderjit 2005; Kaur et al. 2009; Inderjit et al. 2010). While the effect of frozen storage on the microbial

populations of the *P. australis* soil used in this study is unknown, the soil microbial community present clearly was capable of rapidly degrading very large doses of gallic acid. The increased persistence of gallic acid at the largest applied rate (20,000 $\mu\text{g g}^{-1}$) may be due to general antimicrobial effects of this very high dose or to the formation of antimicrobial degradation products.

Phenolic acids have been found to sorb very quickly to soil (Cecchi et al. 2004; Schmidt et al. 2012). Depending on the compound, organic matter and clays can either reversibly or irreversibly sorb phenolic acids. Dalton et al. (1989) showed that the binding of ferulic acid increased with soil organic matter content, and clays had a strong capacity for sorbing ferulic acid. Schmidt et al. (2012) found that sorption of polyphenolic compounds increased with the molecular weight and that hydrophobic compounds were sorbed more strongly than polar ones. Sorption of non-tannin phenolics such as gallic acid was lower than it was for tannins in soils amended with various organic carbon sources (Halvorson et al. 2012). Our results in sterile soil (Figure 3) differ significantly from those of Weidenhamer and Romeo (2004), who found that gallic acid applied at rates of 100-400 $\mu\text{g g}^{-1}$ soil was stable in sterile soils. While we did not test the autoclaved soil in this study to verify that it was sterile, autoclaving greatly extended the persistence of gallic acid in this soil as would be expected if microbes capable of degrading gallic acid were eliminated. The *P. australis* soil used in this study has a high sand content (82%, Table 1) and a much higher organic matter content (11.74%, Table 1) than the Florida scrub soils studied by Weidenhamer and Romeo (no more than 1%), and sorption to organic matter thus seems the most likely explanation for the greater sorption of gallic acid observed here.

311

312 While some researchers have argued that the ability of microorganisms to rapidly degrade many
313 allelochemicals will prevent the manifestation of allelopathy in many field situations (e.g.
314 Schmidt 1988; Kaur et al. 2009), some compounds do show resistance to degradation (Sosa et al.
315 2010) and others degrade to compounds that are more resistant to breakdown, such as the
316 formation of the more toxic and more persistent 2-amino-3*H*-phenoxazin-3-one (APO) from
317 microbial degradation of benzoxazinones released by roots of cereals such as rye (Krogh et al.
318 2006). In addition, it has recently been shown that both polar and non-polar phytotoxins can be
319 transported through soil via the common mycorrhizal network of fungal mycelia that permeate
320 soil, thus providing a potential conduit for these compounds away from microbial hot spots in the
321 rhizosphere (Barto et al. 2011). Other plant systems may continuously produce and release large
322 amounts of allelochemicals over time, regardless of whether the compounds persist in the soil
323 environment for long periods of time (Bertin et al. 2009). However, our studies suggest that it is
324 highly unlikely for gallic acid to be responsible for the allelopathic effects of invasive *P.*
325 *australis*, given the absence of significant levels of gallic acid and gallotannins in the suspected
326 source plant, and the absence of gallic acid in the *P. australis* soils that were tested. Previous
327 bioassay work with gallic acid suggests that persistent degradation products with residual
328 phytotoxicity may be formed from gallic acid (Weidenhamer et al. 1989), but this would require
329 regular inputs of significant amounts of gallic acid into the soil and this does not appear to be the
330 case. Our detection limits for gallic acid in soil were less than 0.5 $\mu\text{g g}^{-1}$, compared to a lowest
331 concentration of more than approximately 70 $\mu\text{mol g}^{-1}$ in fresh soil (approximately 12,000 μg
332 g^{-1}) beneath natural stands of invasive *P. australis* reported by Rudrappa et al. (2007). Bains et
333 al. (2009) reported soil concentrations of 7.2-22.5 $\mu\text{mol gallic acid g}^{-1}$ (= 1225-3828 $\mu\text{g gallic}$

acid g⁻¹) in the rhizosphere of invasive *P. australis* growing in the Delmarva peninsula (USA). Rudrappa et al. (2007) and Bains et al. (2009) used 2.5 mM potassium phosphate buffer (pH = 2.5) for their soil extractions. While different extraction methods for soil phenolics may recover some reversibly bound material (Schmidt et al. 2012; Blair et al. 2005, 2009), the differences here are orders of magnitude, and in addition we have found no evidence of significant gallic acid production by the supposed source plant. We note that neither Rudrappa et al. (2007) nor Bains et al. (2009) report data on soil blank extractions or recovery of gallic acid from spiked soil, both of which are crucial for investigations of potential allelopathic activity (Blair et al. 2009). Bains et al. (2009) reported concentrations of gallic acid that were as much as 10-fold less than those seen by Rudrappa et al. (2007), but this difference was not noted as significant and no data on the seasonal or other dynamics of gallic acid concentrations over time have subsequently been reported by these authors.

Gallic acid content of *P. australis* leaves and rhizomes. Bains et al. (2009) reported that rhizomes of invasive *P. australis* contained 12.2-48.8 µmol gallic acid g⁻¹ FW (= 2075-8302 µg gallic acid g⁻¹ FW), and some 194-265 µmol gallotannin g⁻¹ FW. Gallotannin was quantified using commercial tannic acid as a standard, which is a mixture of condensed and/or hydrolysable tannins obtained from the source plant(s) for the extract (Hagerman et al. 1992; Kinraide and Hagerman 2010). Typically, depending on the content of these tannins, 1 µmol of tannin will yield 4-12 µmol of gallic acid (Kinraide and Hagerman 2010). Assuming a conservative yield of 4 µmol of gallic acid per µmol tannin, 194-265 µmol gallotannin g⁻¹ FW could yield 132-180 mg gallic acid g⁻¹ FW. Combined with the reported amounts of free gallic acid, the expected minimum total extractable gallic acid content in invasive *P. australis* rhizomes is 134-188 mg

gallic acid g⁻¹ FW (equivalent to 13.4-18.8 weight %). By contrast, we found only trace to undetectable amounts of gallic acid in our samples (Table 4). Our detection limits were less than 20 µg gallic acid g⁻¹ FW in the rhizome samples tested, which is approximately 0.015% of the minimum amount of gallic acid expected.

Our results agree with those of Hendricks et al. (2011), who did not quantify gallic acid specifically, but reported low amounts (approximately 0.12% dry mass) of total phenolics in various extracts of *P. australis*. Hendricks et al. (2011) isolated one or more moderately polar compounds that deterred grazing by the abundant salt marsh herbivore *Littoraria irrorata*. The very low concentrations of phenolics isolated from *P. australis* tissues in their study suggested that these are not the source of herbivore deterrence. The thin layer chromatography characteristics of the isolated defense compound(s) were also not consistent with gallic acid. The results from Hendricks et al. (2011) and our current data both suggest that compounds other than gallic acid are involved in the chemical defense of *P. australis* against herbivores and neighboring plants.

Other chemical constituents of *P. australis*. Among the compounds reported from *Phragmites communis* (an early name for *P. australis*, see Mal and Narine (2004)) rhizomes are the anti-algal compound ethyl acetoacetate (Li and Hu 2005) and methyl gallate (Choi et al. 2009). Rudrappa et al. (2007) reported that they were unable to find evidence of ethyl acetoacetate in their root secretion abstracts, and we also found no evidence of this compound by GC-MS analysis of methanol extracts of *P. australis* rhizomes (data not shown). The presence of methyl gallate may

explain the trace amounts of gallic acid present in some rhizome and foliar extracts. We have not yet tested for possible allelopathic effects of other compounds in *P. australis* rhizome and foliar extracts, and so cannot comment on whether any of these may play an allelopathic role for this species. Thus far, 5-HMF is the only major component of the *P. australis* rhizome and leaf extracts that has been characterized. Further investigation of the phytochemistry of *P. australis* is continuing.

Implications of this work for the hypothesis of gallic acid-mediated allelopathy as the mechanism for invasive success of P. australis. Our findings show: (1) Free gallic acid was not detected in *P. australis* soil from North Carolina, in contrast to previous reports of persistent levels of gallic acid in excess of 1000 $\mu\text{g g}^{-1}$ in *P. australis* soils (Bains et al. 2009) to more than 10,000 $\mu\text{g g}^{-1}$ (Rudrappa et al. 2007). (2) Gallic acid rapidly degrades in non-sterile soil, even at very high doses, in contrast to the reported persistence of this compound in soil (Rudrappa et al. 2007; Bains et al. 2009). (3) The rapid degradation of gallic acid in non-sterile soil implies that high amounts of gallic acid could therefore be maintained in *P. australis* soil only through the regular influx of large quantities of gallic acid. However, our results show that only trace amounts of gallic acid were found in samples of rhizomes and foliage of four different *P. australis* populations. Bains et al. (2009) reported very high levels of gallotannins in addition to the free gallic acid present in the plant. Our hydrolysis and extraction procedure was validated using known samples of gallic acid to verify its stability and methyl gallate as a proxy for gallotannins to verify the effectiveness of the acid hydrolysis procedure to cleave bound forms of

gallic acid. The amounts of gallic acid which we found in the *P. australis* rhizomes we tested were less than 0.015% of the minimum amount expected based on previous reports. All four populations sampled have been verified to be of the invasive Eurasian genotype previously identified as possessing both high free gallic acid and gallotannin content (Rudrappa et al. 2007, 2009; Rudrappa and Bais 2008; Bains et al. 2009). (4) Most investigations of the toxicity of gallic acid indicate that it is not particularly toxic to other plants. While Rudrappa et al. (2007) reported that gallic acid inhibited the growth of *Spartina alterniflora* at concentrations of 0.05-0.15 mM, other data in the literature indicate that gallic acid requires higher concentrations to exert toxic effects. Reigosa et al. (1999) tested the effects of six phenolic acids including gallic acid on the germination and growth of six weed species. Concentrations of 10, 1, 0.1 and 0.01 mM were used, and in only one case did gallic acid show any inhibitory activity below a concentration of 10 mM. Chung et al. (2002) also found gallic acid to be less active than several other phenolic acids tested in bioassays against barnyardgrass. Our data indicate that gallic acid concentrations in soil will not reach even the 0.05 mM level at which Rudrappa et al. (2007) reported effects against *Spartina alterniflora* and other species.

In summary, our results do not support the reported role for gallic acid in the invasive success of *P. australis*. Because we have not examined plant populations and soils sampled in previous studies by Rudrappa, Bains and coauthors, we cannot rule out the occurrence of high concentrations of free gallic acid and gallotannins in the populations they tested. Gallotannin concentrations are known to vary seasonally (Salminen et al. 2001), and in response to various environmental factors (Laponen et al. 1998; Agrell et al. 2000; Gross 2003), but we are not aware of any reports showing intra-species variation in gallotannin concentration of the

magnitude which would explain the difference between our results and those of Bains et al. (2009). Our plant populations were located in both freshwater and brackish wetlands, but it is unclear whether the Delaware and Georgia sampling sites of Rudrappa et al. (2007) and Bains et al. (2009) were freshwater, brackish, or marine wetlands. However, experiments reported by these authors included root exudates collected in hydroponic culture and plants grown in pots, with no indication of any differences in chemistry based on water salinity. Saltonstall (2003) notes that the invasive *P. australis* genotype has spread along the North American Atlantic coast as well as inland, suggesting that a common mechanism is responsible for the invasive success of this genotype independent of wetland type. Based on the nearly complete absence of gallic acid in the four populations of invasive genotype that we sampled, we conclude that gallic acid cannot be a primary, general explanation for the invasive success of *P. australis* in North America.

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REFERENCES

- AGRELL, J., McDONALD, E. P., and LINDROTH, R. L. 2000. Effects of CO₂ and light on tree phytochemistry and insect performance. *OIKOS* 88: 259-272.
- BAINS, G., KUMAR, A. S., RUDRAPPA, T., ALFF, E., HANSON, T. E., and BAIS, H. P. 2009. Native plant and microbial contributions to a negative plant-plant interaction. *Plant Physiol.* 151:1699-1700.
- BARTO, E.K., HILKER, M., MÜLLER, F., MOHNEY, B. K., WEIDENHAMER, J. D., and RILLIG, M. C. 2011. The fungal fast lane: Common mycorrhizal networks extend bioactive zones of allelochemicals in soils. *PLoS ONE* 6: e27195.
- BERTIN, C., HARMON, R., AKAOGI, M., WEIDENHAMER, J. D., and WESTON, L. A. 2009. Assessment of the phytotoxic potential of *m*-tyrosine in laboratory soil bioassays. *J. Chem. Ecol.* 35: 1288-1294.
- BLAIR, A. C., HANSON, B. D., BRUNK, G. R., MARRS, R. A., WESTRA, P., NISSEN, S. J., and HUFBAUER, R. A. 2005. New techniques and findings in the study of a candidate allelochemical implicated in invasion success. *Ecol. Lett.* 8: 1039-1047.
- BLAIR, A. C., WESTON, L. A., NISSEN, S. J., BRUNK, G. R., and HUFBAUER, R. A. 2009. The importance of analytical techniques in allelopathy studies with the reported allelochemical catechin as an example. *Biol Invasions* 11: 325-332.
- BLUM, U. 1998. Effects of microbial utilization of phenolic acids and their phenolic acid breakdown products on allelopathic interactions. *J. Chem. Ecol.* 24: 685-708.
- BLUM, U., STAMAN, K. L., FLINT, L. J., and SHAFER, S. R. 2000. Induction and/or selection of phenolic acids-utilizing bulk-soil and rhizosphere bacteria and their influence on phenolic acid phytotoxicity. *J. Chem. Ecol.* 26: 2059-2078.
- CALLAWAY, R.M. and RIDENOUR, W.M. 2004. Novel weapons: invasive success and the evolution of increased competitive ability. *Front. Ecol. Environ.* 2: 436-443.
- CECCHI, A., KOSKINEN, W., CHENG, H., and HAIDER, K. 2004. Sorption-desorption of phenolic acids as affected by soil properties. *Biol. Fertility Soils* 39: 235-242.
- CHOI, S.-E., YOON, J.-H., CHOI, H.-K., and LEE, M.-W. 2009. Phenolic compounds from the root of *Phragmites communis*. *Chem. Nat. Compd.* 45: 893-895.
- CHUNG, I. M., KIM, K. H., AHN, J. K., CHUN, S. C., KIM, C. S., KIM, J. T., and KIM, S. H. 2002. Screening of allelochemicals on barnyardgrass (*Echinochloa crus-galli*) and identification of potentially allelopathic compounds from rice (*Oryza sativa*) variety hull extracts. *Crop Prot.* 21: 913-920.

480 DALTON, B. R., BLUM, U., and WEED, S. B. 1989. Plant phenolic acids in soils: Sorption of
 481 ferulic acid by soil and soil components sterilized by different techniques. *Soil Biol. Biochem.*
 482 21: 1011-1018.

483 DOYLE, J. J. and DOYLE, J. L. 1987. A rapid DNA isolation procedure for small quantities of
 484 fresh leaf tissue. *Phytochem. Bull.* 19: 11-15.

485 GASPAR, E. M. S. M. and LUCENA, A. F. F. 2009. Improved HPLC methodology for food
 486 control – furfurals and patulin as markers of quality. *Food Chem.* 114: 1576-1582.

487 GROSS, E. M. 2003. Differential response of tellimagrandin II and total bioactive hydrolysable
 488 tannins in an aquatic angiosperm to changes in light and nitrogen. *OIKOS* 103: 497-504.

489 HAGERMAN, A. E., ROBBINS, C. T., WEERASURIYA, Y., WILSON, T. C., and MCARTHUR, C. 1992.
 490 Tannin chemistry in relation to digestion. *J. Range Manage* 45:57-62.
 491

492 HALVORSON, J. J., GOLLANY, H. T., KENNEDY, A. C., HAGERMAN, A. E., GONZALEZ, J. M., and
 493 WUEST, S. B. 2012. Sorption of tannin and related phenolic compounds and effects on
 494 extraction of soluble-N in soil amended with several carbon sources. *Agriculture* 2: 52-72.

495 HARBORNE, J. B. 1984. *Phytochemical Methods: A Guide to Modern Methods of Plant*
 496 *Analysis*. 2nd ed, Chapman and Hall, London.

497 HENDRICKS, L. G., MOSSOP, H. E., and KICKLIGHTER, C. E. 2011. Palatability and chemical
 498 defense of *Phragmites australis* to the marsh periwinkle snail *Littoraria irrorata*. *J. Chem. Ecol.*
 499 37: 838–45.

500INDERJIT. 2005. Soil microorganisms: An important determinant of allelopathic activity. *Plant*
 501 *Soil* 274: 227-236.

502INDERJIT, BAJPAI, D., and RAJESWARI, M. S. 2010. Interaction of 8-hydroxyquinoline with soil
 503 environment mediates its ecological function. *PLoS ONE* 5:e12852.

504KAUR, H., KAUR, R., KAUR, S., BALDWIN, I. T., INDERJIT. 2009. Taking ecological function
 505 seriously: Soil microbial communities can obviate allelopathic effects of released metabolites.
 506 *PLoS ONE* 4: e4700.

507KINRAIDE, T. B. and HAGERMAN, A. E. 2010. Interactive intoxicating and ameliorating effects
 508 of tannic acid, aluminum (Al³⁺), copper (Cu²⁺), and selenate (SeO₄²⁻) in wheat roots: a
 509 descriptive and mathematical assessment. *Physiol. Plantarum* 139: 68–79.
 510

511KROGH, S. S., MENSZ, S. J. M., NIELSEN, S. T., MORTENSEN, A. G., CHRISTOPHERSEN, C., and
 512 FOMSGAARD, I. S. 2006. Fate of benzoxazinone allelochemicals in soil after incorporation of
 513 wheat and rye sprouts. *J. Agric. Food Chem.* 54: 1064-1074.

514LAPONEN, J., OSSIPOV, V., LEMPA, K., HAUKIOJA, E., and PIHLAJA, K. 1998. Concentrations and
 515 among-compound correlations of individual phenolics in white birch leaves under air pollution
 516 stress. *Chemosphere* 37: 1445-1456.

517 LI, F.-M. and HU, H.-Y. 2005. Isolation and characterization of a novel antialgal allelochemical
518 from *Phragmites communis*. *Appl. Environ. Microbiol.* 11: 6545-6553.

519 MAL, T. K. and NARINE, L. 2004. The biology of Canadian weeds. 129. *Phragmites*
520 *australis* (Cav.) Trin. ex Steud. *Can. J. Plant Sci.* 84: 365-396.

521 RAMÍREZ-JIMÉNEZ, A., GARCÍA-VILLANOVA, B., GUERRA-HERNÁNDEZ, E. 2000.
522 Hydroxymethylfurfural and methylfurfural content of selected bakery products. *Food. Res. Int.*
523 33: 833-838.

524 REIGOSA, M. J., SOUTO, X. C., and GONZÁLEZ, L. 1999. Effect of phenolic compounds on the
525 germination of six weeds species. *Plant Growth Regul.* 28: 83-88.

526 RUDRAPPA, T. and BAIS, H. P. 2008. Genetics, novel weapons and rhizospheric microcosmal
527 signaling in the invasion of *Phragmites australis*. *Plant Signaling and Behavior* 3: 1-5.

528 RUDRAPPA, T., BONSALE, J., GALLAGHER, J. L., SELISKAR, D. M., and BAIS, H. P. 2007. Root-
529 secreted allelochemical in the noxious weed *Phragmites australis* deploys a reactive oxygen
530 species response and microtubule assembly disruption to execute rhizotoxicity. *J. Chem. Ecol.*
531 33:1898-1918.

532 RUDRAPPA, T., CHOI, Y. S., LEVIA, D. F., LEGATES, D. R., LEE, K. H., and BAIS, H. P. 2009.
533 *Phragmites australis* root secreted phytotoxin undergoes photo-degradation to execute severe
534 phytotoxicity. *Plant Signaling and Behavior* 4: 506-513.

535 SALMINEN, J.-P., OSSIPOV, V., HAUKIOJA, E., and PIHLAJA, K. 2001. Seasonal variation in the
536 content of hydrolysable tannins in leaves of *Betula pubescens*. *Phytochemistry* 57: 15-22.

537 SALTONSTALL, K. 2002. Cryptic invasion by a non-native genotype of the common reed,
538 *Phragmites australis*, into North America. *Proc. Natl. Acad. Sci. USA* 99:2445-2449.

539 SALTONSTALL, K. 2003. A rapid method for identifying the origin of North American
540 *Phragmites* populations using RFLP analysis. *Wetlands* 23:1043-1047.

541 SCHMIDT, M. A., HALVORSON, J. J., GONZALEZ, J. M., and HAGERMAN, A. E. 2012. Kinetics
542 and binding capacity of six soils for structurally defined hydrolysable and condensed tannins and
543 related phenols. *J. Soils Sediments* 12: 366-375.

544 SCHMIDT, S. K. 1988. Degradation of juglone by soil bacteria. *J. Chem. Ecol.* 14: 1561-1571.

545 SOSA, T., VALARES, C., ALÍAS, J. C., and LOBÓN, N. C. 2010. Persistence of flavonoids in *Cistus*
546 *ladanifer* soils. *Plant Soil* 337: 51-63.

547 WEIDENHAMER, J. D., HARTNETT, D. C., and ROMEO, J. T. 1989. Density-dependent
548 phytotoxicity: Distinguishing resource competition and allelopathic interference in plants. *J.*
549 *App. Ecol.* 26: 613-624.

550 WEIDENHAMER, J. D. and ROMEO, J. T. 1989. Allelopathic properties of *Polygonella myriophylla*:
551 Field evidence and bioassays. *J. Chem. Ecol.* 15:1957-1969.

552 WEIDENHAMER, J. D. and ROMEO, J. T. 2004. Allelochemicals of *Polygonella myriophylla*:
553 Chemistry and soil degradation. *J. Chem. Ecol.* 30:1061-1078.

554 WEIDENHAMER, J. D. and ROMEO, J. T. 2005. Allelopathy as a mechanism for resisting invasion:
555 The case of *Polygonella myriophylla*, pp. 167-177, in Inderjit (ed.). *Invasive Plants: Ecological*
556 *and Agricultural Aspects*. Birkhauser Verlag, Switzerland.

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TABLE 1. PHYSICAL AND CHEMICAL CHARACTERISTICS OF *P. AUSTRALIS* SOIL

Parameter	Value
pH	6.96
P, $\mu\text{g g}^{-1}$	54.8
K, $\mu\text{g g}^{-1}$	156
Ca, $\mu\text{g g}^{-1}$	2812
Mg, $\mu\text{g g}^{-1}$	268
Cation exchange capacity, meq 100 g^{-1}	16.7
Organic matter, %	11.74
Sand content, %	82%

TABLE 2. DEGRADATION RATE CONSTANT, K, AND HALF-LIFE, $T_{1/2}$, FOR
BREAKDOWN OF GALLIC ACID IN NONSTERILE SOIL

Concentration ($\mu\text{g g}^{-1}$)	r^2	k ($\text{day}^{-1} \pm 1 \text{ SE}$)	$t_{1/2}$ (days $\pm 1 \text{ SE}$)
500	0.95	2.17 (± 0.13)	0.32 (± 0.02)
1000	0.89	2.16 (± 0.19)	0.32 (± 0.03)
5000	0.91	2.29 (± 0.15)	0.30 (± 0.02)
10,000	0.95	0.76 (± 0.04)	0.91 (± 0.04)
20,000	0.80	0.22 (± 0.02)	3.13 (± 0.27)

TABLE 3. STABILITY AND RECOVERY OF GALLIC ACID FROM ACID HYDROLYSIS

PROCEDURE

Amount	Gallic acid (% of expected \pm S.D.)	Methyl gallate (% of expected \pm S.D.)
2 mg	98.6 \pm 1.2	84.0 \pm 8.9
10 mg	83.0 \pm 2.3	73.3 \pm 1.0
25 mg	72.4 \pm 5.4	70.6 \pm 1.8

TABLE 4. GALLIC ACID CONTENT OF RHIZOMES AND LEAVES FROM FOUR
POPULATIONS OF *P. AUSTRALIS*

Population ^a	Rhizome sample 1		Rhizome sample 2	
	$\mu\text{g g}^{-1}$ FW found ^b	detection limit ^c	$\mu\text{g g}^{-1}$ FW found	detection limit
OH-1	ND	10	ND	6.3
OH-2	ND	11	ND	7.8
NC-1	ND	17	TR	13
NC-2	TR	7.4	TR	5.2
	Foliage sample 1		Foliage sample 2	
	$\mu\text{g g}^{-1}$ FW found	detection limit	$\mu\text{g g}^{-1}$ FW found	detection limit
OH-1	ND	14	TR	14.6
OH-2	20	18	28	16
NC-1	TR	14	TR	13.1
NC-2	TR	13	TR	17.6

^aOH = Ohio; NC = North Carolina. See methods for information on sample collection.

^bND = not detected; TR = trace level (below quantitation limit but definite peak at retention time for gallic acid)

^cDetection limits calculated assuming quantitation limit of 0.25 mg l^{-1} and 70% recovery of gallic acid through hydrolysis and extraction procedure.

FIGURE LEGENDS

FIG. 1. Degradation of exogenously applied gallic acid in nonsterile *P. australis* soil.

Standard errors for all data points in Fig. 1A were less than 3% and are not shown. In Fig. 1B, bars indicate standard error based on triplicate determinations.

FIG. 2. Gallic acid breakdown in soil monitored by HPLC. Chromatograms are based on absorbance at 272 nm, and show unidentified microbial breakdown products (marked by *) observed after addition of gallic acid ($500\text{ }\mu\text{g g}^{-1}$) to nonsterile *P. australis* soil. (A) Initial extraction immediately after spiking soil; (B) 0.5 days; and (C) 1.5 days.

FIG. 3. Stability of exogenously applied gallic acid in sterile *P. australis* soil. Standard errors for all data points in Fig. 3 were less than 4% and are not shown.

FIG. 4. RFLP analysis of Ohio and North Carolina *P. australis* samples. Shown are ethidium bromide stained 3% agarose gels containing PCR products of two chloroplast regions, *rbcL* and *trnLb*, from the four collection sites used in this study. The *rbcL* region of invasive strains contains an HhaI restriction enzyme recognition site while the *trnLb* region of native strains contains an RsaI restriction enzyme site. The presence of two reduced band sizes in the HhaI digested *rbcL* products indicates that all four collection sites contained invasive strains.

FIG. 5. Separation of major components of *P. australis* rhizomes and foliage.

Chromatograms are based on absorbance at 272 nm and show separation of major components of rhizomes and foliage: (A) Ohio population 2 rhizome; (B) Ohio population 2 foliage (inset shows apparent gallic acid peak); (C) North Carolina population 1 rhizome; (D) North Carolina population 1 foliage. The peak eluting at 15.9 min (marked by an asterisk in each chromatogram below) was identified as an artifact of the acid hydrolysis extraction procedure (5-hydroxymethylfurfural). The other compounds in the extract are yet to be identified. The difference in retention time for gallic acid compared to Fig. 2 is due to aging of the HPLC column.

Supplemental FIG. 1. Mass spectrum of 5-hydroxymethylfurfural isolated from *P. australis* extract (A) compared to NIST library spectrum (B).

Supplemental FIG. 2. Mass spectrum of suspected 2-methoxy-4-vinylphenol isolated from *P. australis* extract (A) compared to NIST library spectrum (B).

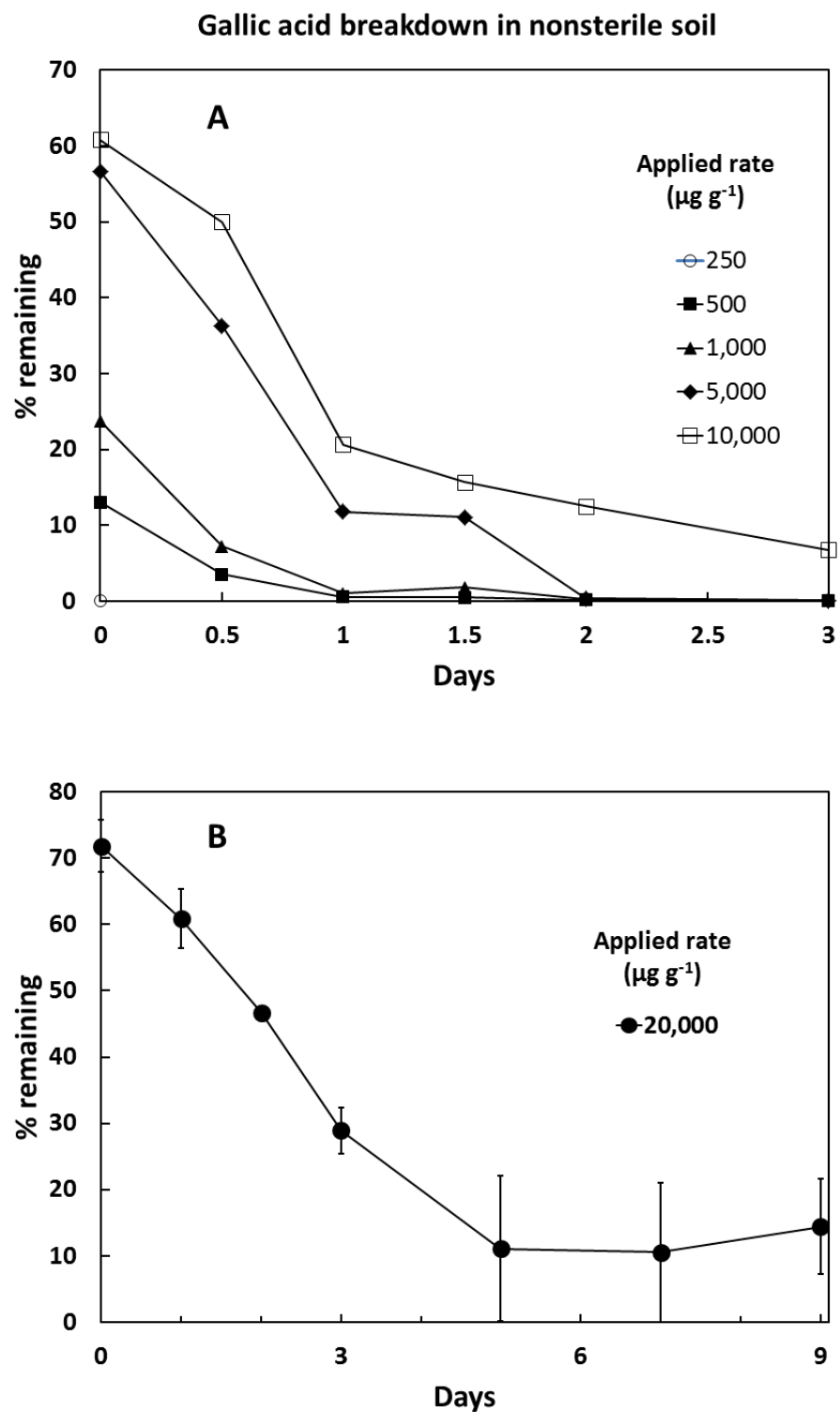


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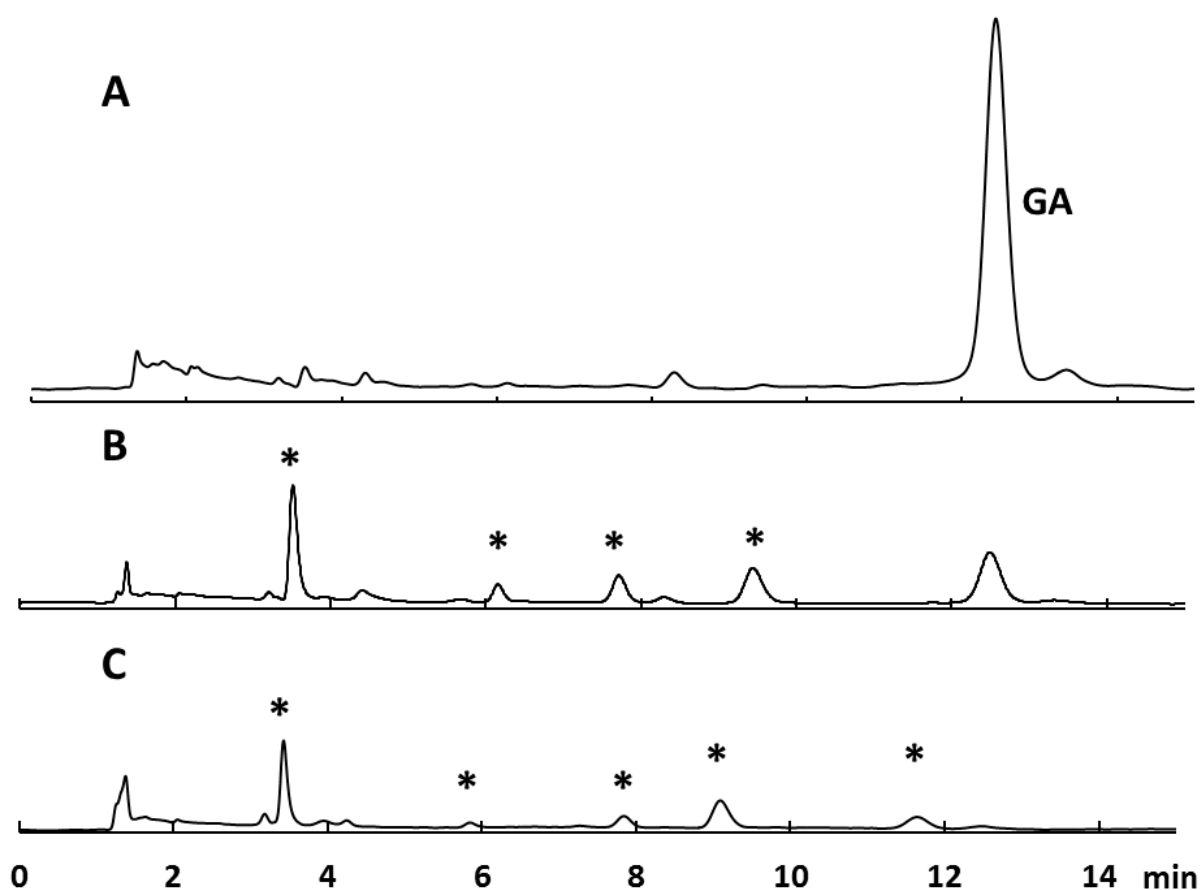


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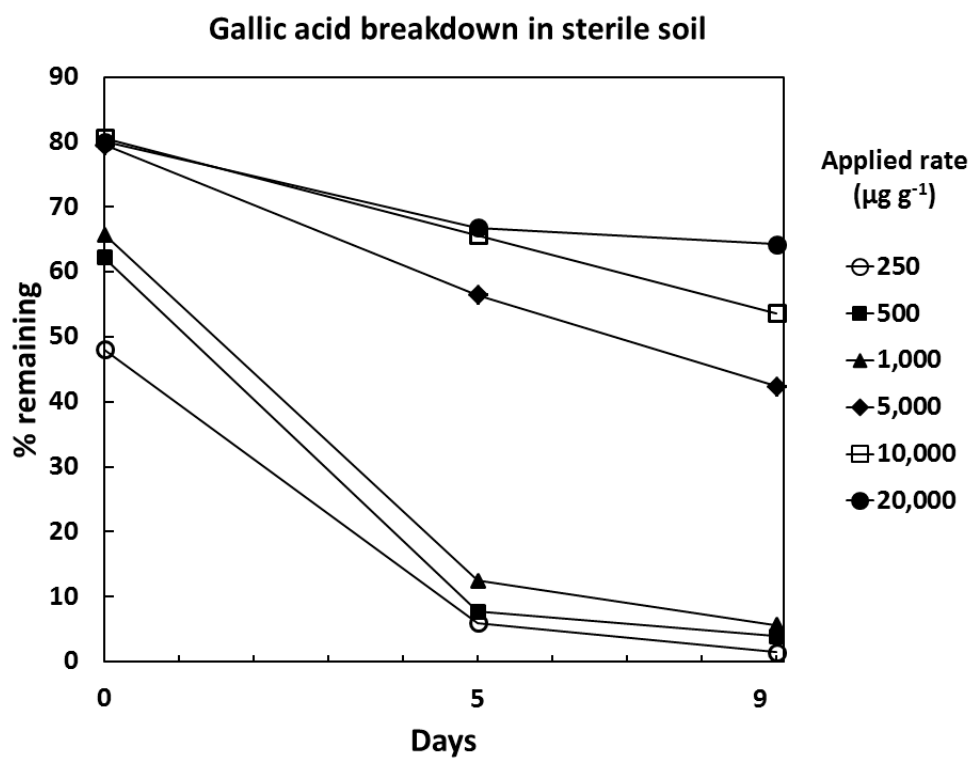


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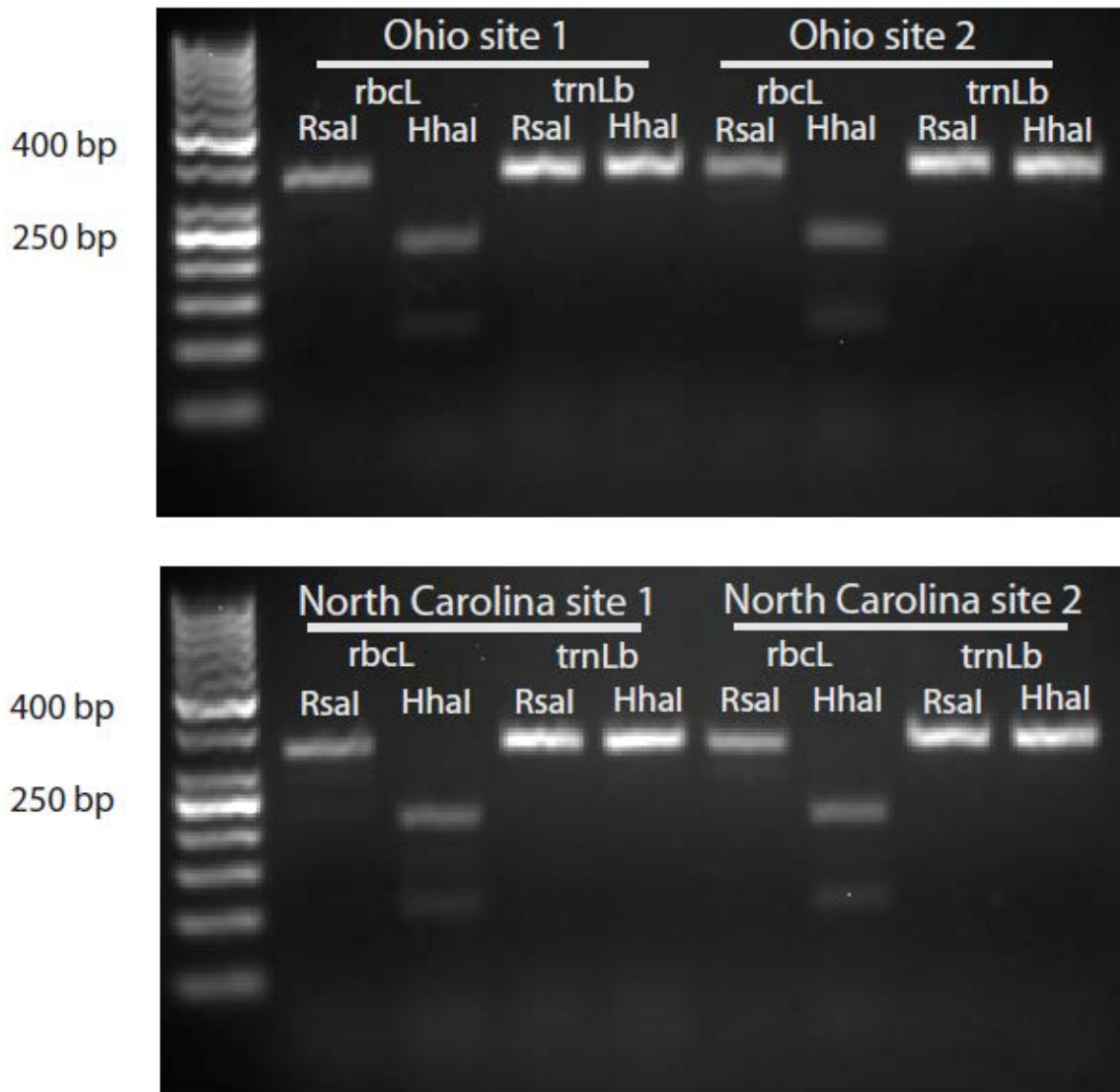


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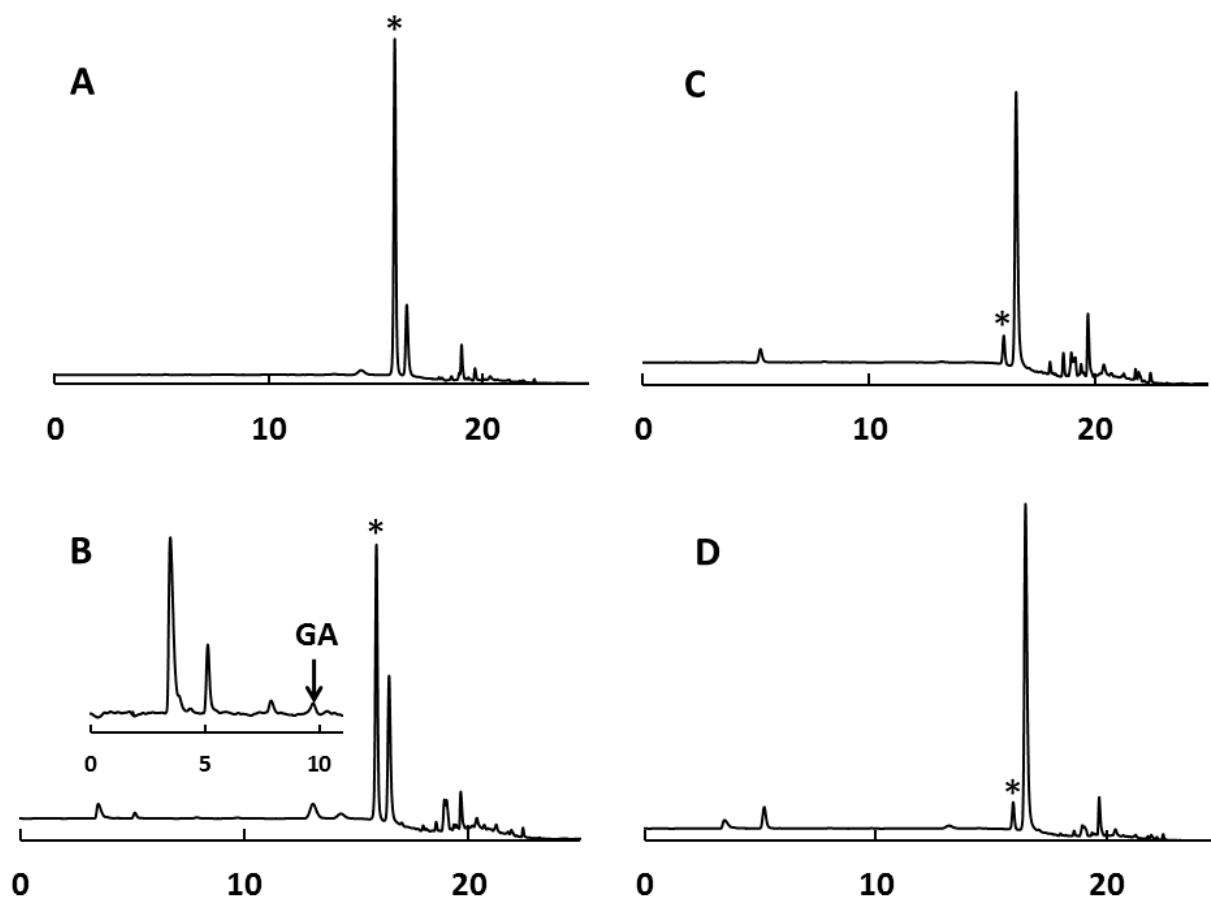
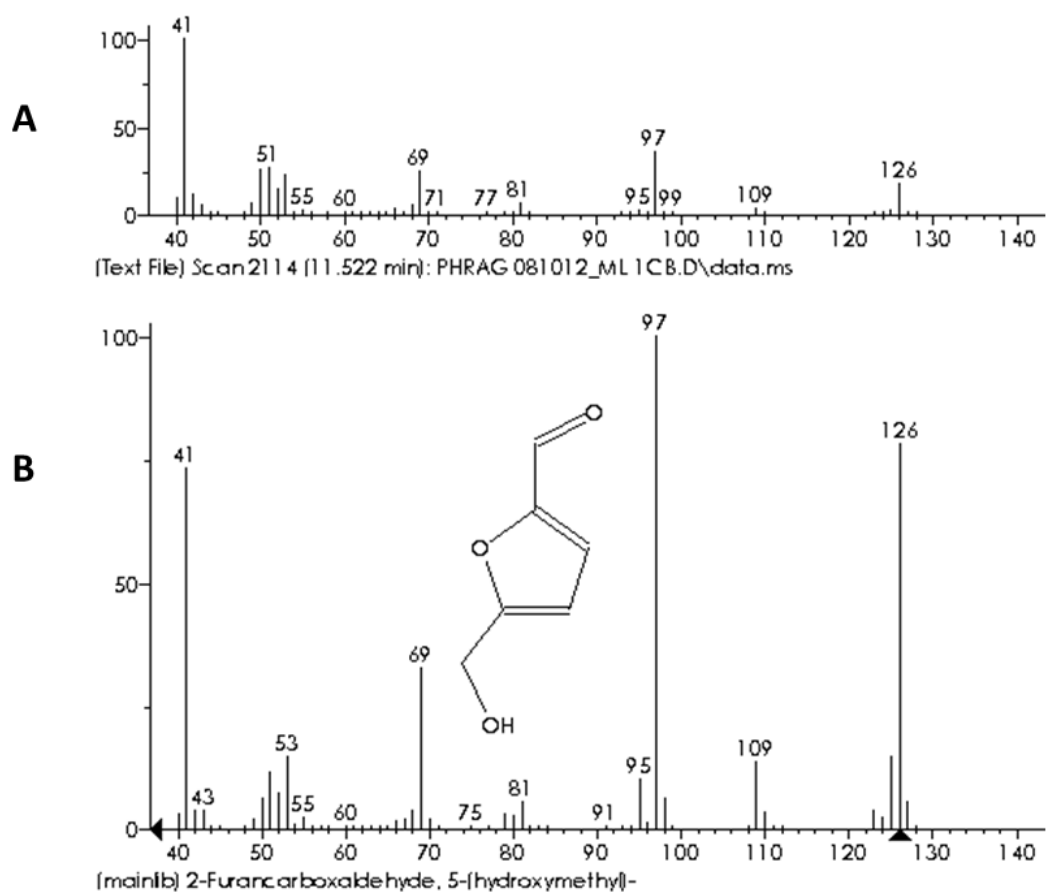
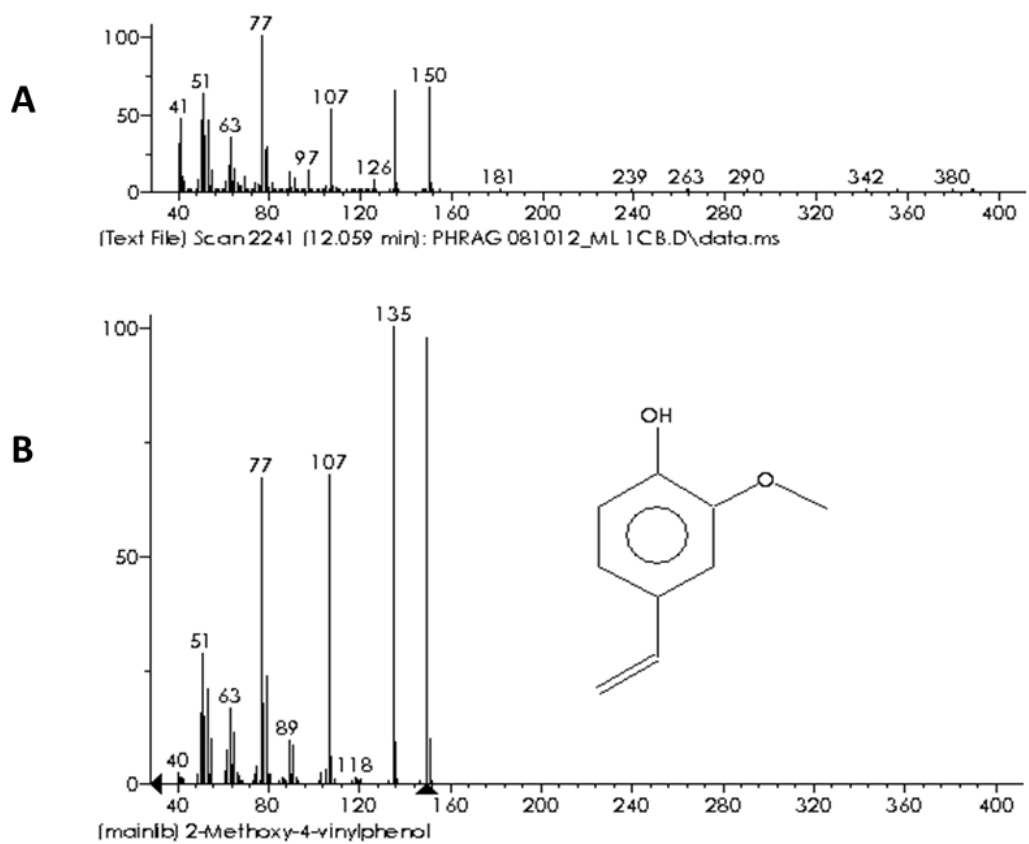


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